

Marine Bioregions in Australia – Numerical Analysis of Provinces on the East Coast



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the Degree of Master of Research in Biological Sciences*

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Declaration

I wish to acknowledge the following assistance in the research detailed in this report:

N/A

All other research described in this report is my own original work.

Matthew Roy Kerr, MRes, DIC, BSc

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Abstract

Bioregions, distinct ecological assemblages of flora and fauna, have long been used to aid conservation decisions. Bioregions containing at-risk taxa have routinely been recommended for assignment as Marine Protected Areas (MPA). However, the number, location and extent of bioregions in Australia has been heavily debated over the previous few decades. The current consensus, based on coastal fishes, does not take into account the diverse benthic taxa present.

A large dataset of Australian east coast molluscs, assembled from museum collection records of the last 100 years, was analysed to identify bioregions. Latitudinal bands were analysed with an array of methods previously utilised to identify bioregions elsewhere in the world. The number and size of provinces from each method were compared to form a consensus.

Faunal turnover was high and consistent across the entire geographic extent, with no distinct separations into bioregions. There was strong support for a distinct Victoria/Tasmania province. Nitrate content, showed the strongest control over southern latitudes and it is likely this province is controlled by nutrient availability.

These results suggest that Australian mollusc distributions are potentially controlled by oceanic currents and nutrient supply from southern waters, with the Bass Strait forming the only distinctive distribution control.

Keywords: Biogeography, Australia, Bioregionalisation, Mollusca, Macroecology

Contents

Acknowledgments	1
Abstract.....	2
Introduction.....	4
<i>Latitudinal biodiversity gradients.....</i>	<i>4</i>
<i>Biological provinces in coastal eastern Australia</i>	<i>7</i>
<i>Project scope.....</i>	<i>8</i>
<i>Taxonomic focus</i>	<i>9</i>
<i>Using natural history collections in ecological studies</i>	<i>10</i>
Methods.....	12
<i>Data</i>	<i>12</i>
<i>Latitudinal richness</i>	<i>15</i>
<i>Biogeographic structure</i>	<i>16</i>
<i>Abiotic variables</i>	<i>18</i>
Results	20
<i>Latitudinal gradients.....</i>	<i>20</i>
<i>Biogeographic structure</i>	<i>22</i>
<i>Relation of biogeographic structure to abiotic variables</i>	<i>28</i>
Discussion.....	33
<i>Interpretation of results</i>	<i>33</i>
<i>Tasmania as a system.....</i>	<i>37</i>
Conclusions	38
<i>Recommendations for further study</i>	<i>39</i>
References	41
Appendices.....	47
Author Guidelines for Journal of Biogeography	48

Introduction

Latitudinal biodiversity gradients

The distinct drop in species diversity from the equator to the poles, termed the latitudinal biodiversity gradient (LBG), is one of the most striking phenomena present in macroecology. As a result, it has been the subject of intense research interest (Willig *et al.*, 2003; Mittelbach *et al.*, 2007). The LBG has been witnessed in almost every part of the tree of life, from mammals (Willig & Lyons, 1998) to marine planktonic bacteria (Fuhrman *et al.*, 2008) and has been shown to have existed since at least the Mesozoic (Cecca, 2002; Fanti & Miyashita, 2009).

The drivers behind the LBG have been explored as early as the 19th Century (Pianka, 1966; Willig, Kaufman, & Stevens, 2003), with several authors conducting large scale reviews into the causes of this distinctive ecological gradient (Rohde, 1992; Blackburn & Gaston, 1996; Roy, Jablonski, & Valentine, 2000; Willig *et al.*, 2003). However, very few drivers are currently agreed upon.

Early explanations focused on the physiology of organisms and how they might not be independent of latitude, notably body size (Blackburn & Gaston, 1996) and geographic range (Stevens, 1989). The controversial "Rapoport's Rule" (Stevens, 1989) holds that the LBG is due to changes in geographic ranges, with organisms living in the tropics having far smaller ones. This theory has widely been put down to a limited pattern as opposed to a rule (Roy, Jablonski, & Valentine, 1994; Gaston, Blackburn, & Spicer, 1998), with no evidence for the effect occurring in latitudes higher than 40°N for North American coastal fishes (Rohde, 1992). Stevens (1989) indeed only accounts for groups of organisms above the tropics, with no data shown for groups above 25°N. The "Rapoport Effect" appears instead to be related to the distribution of oceanic barriers (Roy *et al.*, 1994) and, at least in fishes, to depth range and richness (Smith & Brown, 2002). Body size, too, has widely been related to major oceanic barriers, with

significant relationships between body size and latitude at regional scales and in the tropics, but not on a global scale (Roy & Martien, 2001).

Today, the most widely accepted correlate of latitudinal diversity is energy availability, with sea surface temperature used as a proxy. It exhibits strong correlations with diversity in disparate groups (Roy, Jablonski, Valentine, & Rosenberg, 1998; Roy et al., 2000; Smith & Brown, 2002; Hillebrand, 2004; Jablonski et al., 2013). Evolutionary models show preferential origination of species in the tropics, with a range increase towards the poles (Rohde, 1992; Jablonski, Roy, & Valentine, 2006) and evidenced by steeper gradients in younger taxa (Crame, 2000) and time-area relationships (Fine & Ree, 2006). In the latter case, temperate taxa are shown to be younger and often nested within tropical taxa; diversity in a region correlates with the age of the region (Mittelbach et al., 2007). It has also been shown that fossil taxa preferentially originate in the tropics more so than sampling alone can explain (Jablonski, 1993) and that turnover rates of Recent bivalves were far higher in tropical groups compared to temperate groups (Flessa & Jablonski, 1996).

This correlation is not seen in the southeastern Pacific, where there is a sharp increase in southern latitudes that only corresponds to an increase in shelf area (Valdovinos, Navarrete, & Marquet, 2003) – something that was not found to be significant for the eastern Pacific or western Atlantic (Roy et al., 1998) and may be an effect over a general rule.

Biogeography and biological provinces

The structure of the LBG has also been linked to the presence of biological provinces in North America (Roy et al., 1994), making it imperative to define the number of provinces, the location of their boundaries, and their relationship to diversity patterns. Biological provinces are defined as areas having distinct faunal assemblages contained within dynamic spatial boundaries (Welsh, 1994). Such units have been recognised on several spatial scales, with global scales (Spalding et al., 2007; Oliver & Irwin, 2008) and continental scales (Cook & Auster, 2007;

Last et al., 2010; Narayanaswamy et al., 2010) dominating the macroecological literature.

The exact assignment of biological provinces has been contentious, with many authors assigning provinces based on different criteria: variations in taxonomic group, ordination methodology or spatial scale often result in defining vastly different provinces for the same stretch of land (Briggs, 1995). Traditional assignments have been based on quantifying faunal similarity between latitudinal bands (Valentine, 1966), with binary similarity coefficients being used to recognise tightly clustered groups of similar sites. More recent studies focus on ordination techniques, with it being particularly popular to define province boundaries defined using non-metric multidimensional scaling (Hale, 2010; Williams et al., 2010). In addition to using the broad structure of the LBG, range endpoints – the clustering of range limits for species in a group – have been used to define oceanic boundaries in the North American coastline (Roy et al., 1998; Roy, 2001; MacPherson, 2003) and have shown results similar to Valentine (1966).

Whilst some recent studies have focused on smaller-scale provincial boundaries, either substituting ecological abundance data for genetic connections within a species (e.g. Teske, Sandoval-Castillo, Waters, & Beheregaray, 2017) or focusing on very local abiotic controls on organismal spread (Hale, 2010; Williams et al., 2010), global approaches to biogeography are still commonplace. Global ocean biogeography has been determined by large scale trends in ocean colour, related to plankton abundance and chlorophyll-a concentration (Devred, Sathyendranath, & Platt, 2007a, 2007b; Harvey, Kratzer, & Andersson, 2015) as well as in terms of *dynamic* province boundaries changing due to seasonal effects (Oliver & Irwin, 2008). These approaches, although useful in a global ocean network approach, are often not suitable for looking at small scale processes used in coastal management decisions and discount organism-scale interactions when determining boundaries.

Biological provinces in coastal eastern Australia

Biological provinces have long been used in Australia and beyond for the assignment and assessment of potential marine protected areas (MPAs) (CSIRO, 1996; Kriwoken, 1996; Cook & Auster, 2007; Edgar et al., 2014) and in water zonation for the Exclusive Economic Zone. The separation of waters into regions and reserves placed boundaries between provinces in order to capture features of neighbouring provinces fully (Commonwealth of Australia, 2006; Last et al., 2010); biogeographical splitting remains important in the management of Australian waters.

MPAs currently number well over 250 in Australia and 900 worldwide (McNeill, 1994; Mora et al., 2006; Edgar et al., 2014), with more being added sporadically. The value and usefulness of MPAs as conservation enablers has been controversial (Allison, Lubchenco, & Carr, 1998; Edgar et al., 2014) due to commonality of poor design and lack of proper fishing controls around their borders. However, several reports have also noted the positive effects of MPAs on community structure and resilience (Creighton, Boon, Brookes, & Sheaves, 2015; Mellin, Aaron MacNeil, Cheal, Emslie, & Julian Caley, 2016). Effective MPA design, and the construction of a system of biological provinces to aid this design, is vital to the success of future MPA assignment.

Currently, however, biological provinces in Australia are vastly understudied, with only the occasional extensive analysis having been carried out. An initial study by Bennett and Pope (1952) looked at southern water biogeography in terms of full faunal assemblages. It defined three provinces on the east coast, with tropical provinces taken from Hedley (1904), a broad northern province (“Solanderian”), a south eastern province (“Peronian”) and a separate cool-water province for Victoria and Tasmania (“Maugean”). Knox (1960, 1963) divided the coastline similarly based on physical characteristics, with further support for a Tasmania-Victoria province separate from the rest of Australia again determined by Bennett and Pope (1960). This study is easily the most

comprehensive in its taxonomic scope. However, the methods of analysis are dated.

Wilson and Gillett (1974), however, argue for a much simpler structure comprised of a northern (“tropical”) and southern (“temperate”) province spanning the entire longitudinal extent of Australian waters. Their assessment was based exclusively on gastropod presences. They proposed broad overlap zones stretching from approximately 25.5°S to 37.5°S on the eastern coast, with Tasmania included in the southern geographic zone and not as a separate province. Their conclusions were based purely on observational data, and specifically on shell collection and personal knowledge.

A more extensive study done by CSIRO (1996) has been used as a guide to large-scale Australian provinces (Last et al., 2010). Based on coastal fish distributions, the CSIRO delimited four smaller provinces for demersal fishes on the east coast, with broad overlap zones in between: a Tasmanian province, a Bass Strait Province, a Central Province and a Northern Province. Purely pelagic fishes show a much simpler pattern, with a broad central province, a northern province and Tasmania-Victoria lying as a separate province. Fishes were used as the data for other groups were unreliable. However, their western Australian provinces have been confirmed using megabenthos (Williams et al., 2010).

The east coast of Australia remains unconfirmed both in terms of province boundaries and drivers behind these boundaries. In northwestern Australia, organism dispersal is controlled by trench locations (Last & Séret, 1999) and in the US boundaries are controlled by a variety of factors including temperature (Hale, 2010) and plate tectonics (Gottscho, 2016). These have not yet been related to boundaries on the east coast of Australia.

Project scope

With provinces being so important to conservation efforts and coastal management, a robust system for identifying them is key. This project aims to

elucidate biogeographical provinces on coastal eastern Australia objectively, without making assumptions about previously proposed province boundaries, using a dataset of coastal bivalves and gastropods constructed from museum records across Australia. A large series of data transformations and methodologies will be applied to the dataset.

The newly defined province boundaries are compared to latitudinal trends in biodiversity across the same geographic extent, as well as to abiotic variables that have been shown to control molluscan presence: temperature, salinity, dissolved oxygen content and nutrient availability. Hence, this project not only looks for boundaries, but attempts to identify their consequences and controls.

Taxonomic focus

Molluscs have long been used to study biogeographic patterns (e.g. Coomans, 1962; Valentine, 1966; Wilson & Gillett, 1974; Spalding et al., 2007; Hale, 2010; Itsukushima & Shimatani, 2015) and latitudinal gradients (e.g. Roy et al., 1998; Roy & Martien, 2001; Valdovinos et al., 2003; Jablonski et al., 2006) due not only to their relative ease of collection but also to their lack of motility. Although planktonically dispersed, many adult bivalves and gastropods are benthic and thus remain in one location (Hunt, Maltais, Fugate, & Chant, 2007). Unlike fishes, the usual choice for assigning provinces in Australia, the presence of such organisms in a location is a direct indicator of long-term ecosystem structure.

Molluscan presence has also been linked to ecosystem health (van der Meij, Moolenbeek, & Hoeksema, 2009; Itsukushima & Shimatani, 2015) and may offer insights into anthropogenic effects on ecosystem conditions (Edgar & Barrett, 2000). These properties make them excellent candidates for biogeographic assessment within a conservation framework, especially as they are understudied in terms of extinction risk (Régnier, Fontaine, & Bouchet, 2009).

Using natural history collections in ecological studies

The use of natural history collections (NHCs) in large scale ecological studies has been heavily discussed in recent literature. They are key sources of organismal occurrence data, collected over several hundred years in many cases (Lane, 1996). The deep temporal extent of NHCs makes them especially useful for tracking invasive species (Zalba, Sonaglioni, Compagnoni, & Belenguer, 2000; Graham, Ferrier, Huettman, Moritz, & Peterson, 2004).

There are major issues with NHCs. For example, the differential collection of rare species often can lead to bias in studies that require quantifying ecological abundance (Magurran et al., 2010; Garcillán & Ezcurra, 2011). Likewise, changes in collection methodologies can also create spurious historical trends in relative abundance (Shaffer, Fisher, & Davidson, 1998; Boakes et al., 2010). Apparent latitudinal shifts may also reflect changes in collection effort (Huisman & Millar, 2013), and purported endemism hotspots may reflect sampling artefacts instead of true diversity peaks (Nelson, Ferreira, da Silva, & Kawasaki, 1990). Conversely, museum collections have actually been shown to offer better long term coverage than published literature or using publicly sourced observations (Boakes et al., 2010), and biases may be eliminated by long-term averaging (Lister & Climate Change Research, 2011).

Data for this project is sourced from the Online Zoological Collections of Australian Museums (OZCAM). OZCAM comprises biodiversity information compiled from several museum and university collections. Data accuracy is key to using NHCs for ecological studies (Boakes et al., 2010; Booth, Williams, & Belbin, 2012). OZCAM offers complete consistency in terms of both taxonomic identification and location information, making it preferable for purposes of this study to sources such as the Atlas of Living Australia which includes information from citizen science projects and observations that are not checked by taxonomic experts.

Project Aims

1. To construct a database of molluscan occurrences and ranges for the east coast of Australia, and use it to investigate the key drivers of coastal molluscan latitudinal diversity in terms of abiotic variables.
2. To designate molluscan biogeographical provinces for the east coast of Australia by applying several statistical methods that have been used in the literature to assign provincial boundaries.
3. To investigate what abiotic variables that define the start and end of each province, and to see if they are instead purely controlled by geographic features.

Methods

Data

All gastropod and bivalve records for the east of Australia were downloaded from OZCAM on 8th March 2017, with a latitudinal extent of 10-45°S and a longitudinal extent of 142-155°E (figure 1). In total, seven institutional collections were used (full list in appendix Table 1).

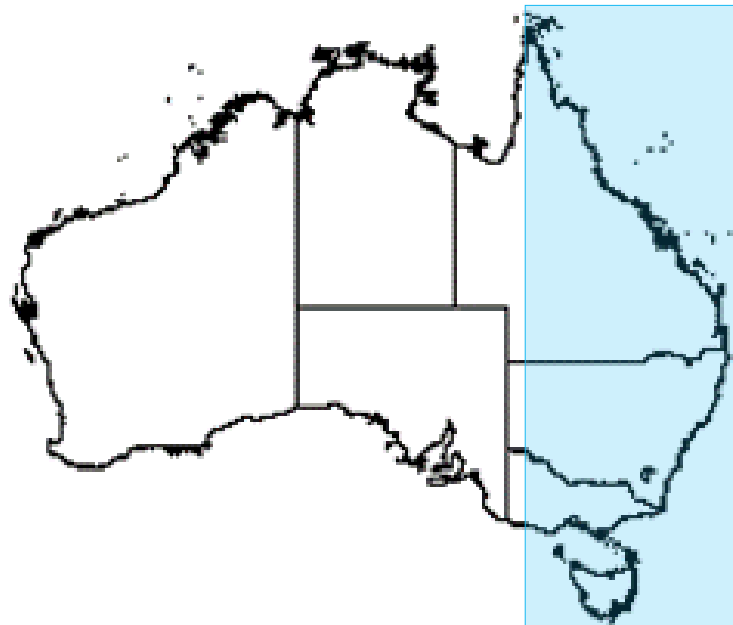


Figure 1: Map showing the geographic extent of records included in the study. Only coastal records from this area were used. Map was drawn using data obtained from Global Administrative Areas spatial database (<http://www.gadm.org>) and visualised in R.

Records with coordinates having a precision of only 0 to 4 decimal places were excluded to ensure correct placement of the records in latitudinal bands (Ponder, Carter, Flemons, & Chapman, 2001; Graham et al., 2004; Boakes et al., 2010). Family- and genus-level assignments were available for all records. However, their utility is debated with respect to resolving regional differences and serving as units of biodiversity (Roy, Jablonski, & Valentine, 1996;

Humphreys & Barraclough, 2014). Any record without a species name was also excluded from the dataset.

Taxonomic checks were carried out on each record using the World Register of Marine Species (WoRMS) web service (<http://www.marinespecies.org/> - Horton et al., 2017) to ensure correct naming of each species, genus and family and to remove junior synonyms. Each taxon was also matched to an environment based on its genus name using the same service. Species that could not be matched, genera matched to an environment other than marine, or genera with unknown affinity were removed from the data. To minimise collecting biases, so were singleton species, i.e., species only occurring at one site, and sites yielding exactly one occurrence.

The final dataset contains 168,601 occurrence records (figure 2) comprising 3796 species, of which gastropods comprise 76% and bivalves 24%. Historically, most of data were collected between 1950 and 2010, with a median year of 1972. Over time, the mean latitudinal position of sampling sites has fluctuated by over 6° of latitude (figure 3). As a result, data were not restricted temporally as to not introduce any latitudinal bias resulting from incomplete sampling in any given period.

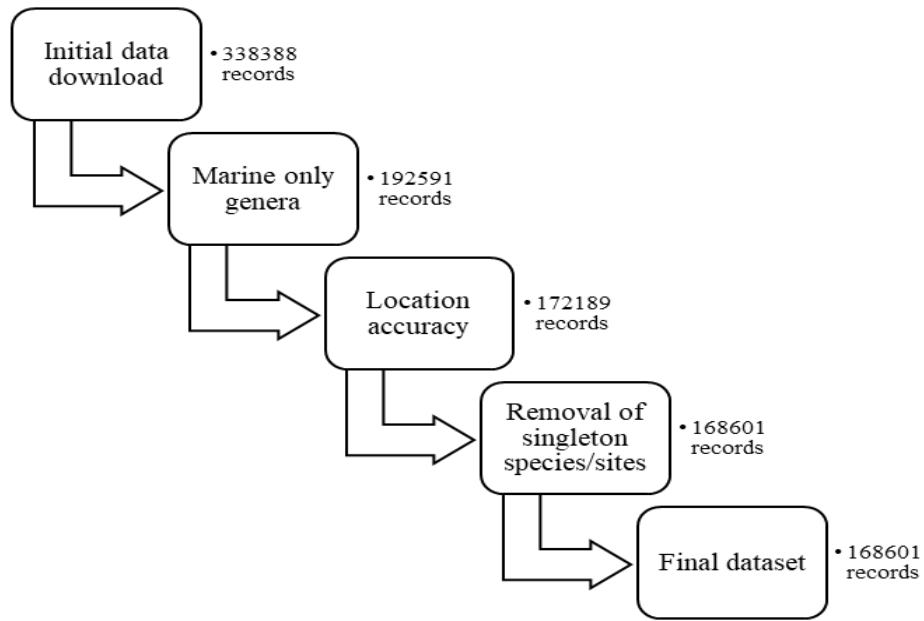


Figure 2: Data loss during cleaning process of the dataset used in this study. Initial download consisted of 338,388 molluscan records, and it was then reduced by removing non-marine data, restricting location information to coordinates of 4 decimal places or higher, and removing singleton species and sites. The final dataset numbered 168,601 records.

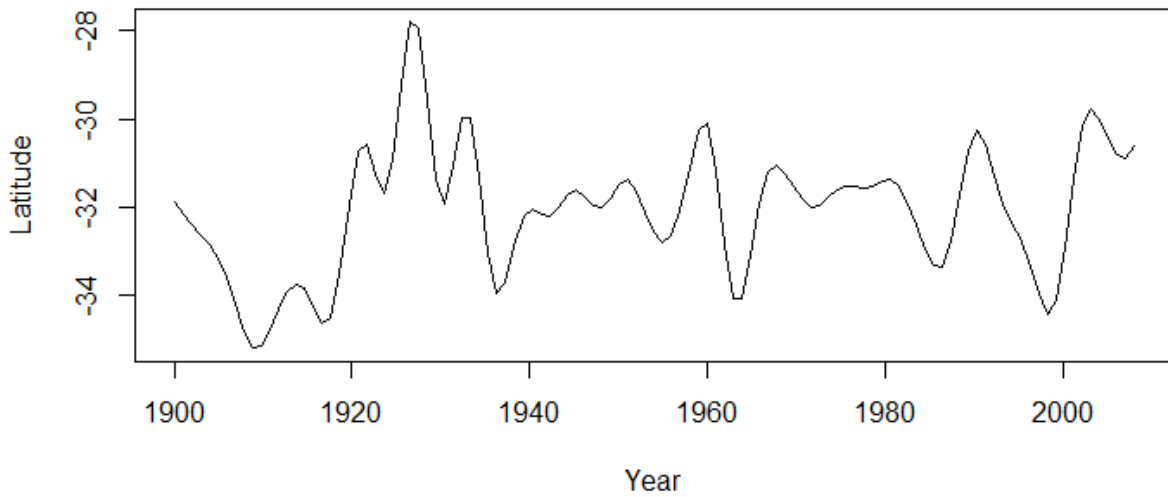


Figure 3: Mean location of sample sites in the dataset. Mean latitude for unique sampling sites were taken every three years for the period 1900 to 2010, which covers 96% of the data records.

Abiotic variables were downloaded from the CSIRO Atlas of Regional Seas (CARS) using the Australian Ocean Data Network portal (<https://portal.aodn.org.au/>). Data was limited to the same geographic extent as the occurrence data. The variables downloaded were temperature, salinity, dissolved oxygen content, phosphate content and nitrate content of the water.

Temperature, dissolved oxygen content, salinity and nutrient content (nitrate and phosphate levels) data were averaged for the years 2007 through 2012. Temperature has been shown to be a good proxy for energy availability (Roy et al., 1998; Roy et al., 2000), while salinity and dissolved oxygen content put physiological and reproductive stress on organisms (Møllergaard & Nielsen, 1995; Chaparro, Cubillos, Montiel, Paschke, & Pechenik, 2008; Segura et al., 2016). Each of these variables can be linked to the formation of provinces and province boundaries, with temperature being a likely explanation for the formation of the latitudinal diversity gradient in some cases (Jablonski et al., 2013). Ocean colour has also been linked to the location of latitudinal provinces globally (Devred et al., 2007a; Oliver & Irwin, 2008) and at regional scales (Ridgway, Dunn, & Wilkin, 2002; Condie & Dunn, 2006; Devred, Sathyendranath, & Platt, 2009; Harvey et al., 2015). It is a function of the amount of organic matter (plant and planktonic) in a system.

Final datasets, both of abiotic variables and species records, are archived and freely available on the Open Science Foundation website (<https://osf.io/twx4m/>).

Latitudinal richness

Records were grouped by latitude, rounded either to the nearest degree or to the nearest 2 degrees. Only 14% of the records contained abundance or count information, so instead, the number of records ('presences') of each species was used as a substitute. A 'range-through' matrix was also created, in which a species was considered present if its latitudinal range crossed a given band (Roy et al.,

1994). Species ranges were taken from the dataset: each location was rounded to the nearest degree, and the most northern and southern degree band containing at least two records was taken as the range limit.

The species richness for each band was estimated using Chao 1, Chao 2 and λ_5 extrapolation methods (Chao, 1984; Chao & Chiu, 2006; Alroy, 2017). Three hundred and fifty records were randomly pulled from each latitudinal band 1000 times, with the geometric mean of all trials taken as the richness for that band. This protocol accounts for sample size variation in the data. In addition, species richness was estimated directly using the ‘range-through’ matrix.

For each band, the mean and median range size of all species it included was also calculated. All statistical analyses were computed in R (R Core Team, 2017). The entire pipeline is available from the author on request.

Biogeographic structure

A partial constrained correspondence analysis (CCA) following the implementation of Legendre and Legendre (2012) was carried out on each version of the dataset using the R-package ‘vegan’ 2.4-3 (Cajo, 1986; Oksanen et al., 2017) to visualise changes in ecological structure across space. A principal coordinates analysis (PCoA, also called classical multidimensional scaling/Torgerson-Gower scaling (Gower, 1966)) was also carried out to complement the CCA. The reason is that CCA can often yield results that are hard to interpret (Legendre & Gallagher, 2001), such as presenting large outliers. All ordination methods depict differences between sites visually, with sites clustering together when they are similar in terms of faunal composition.

An ordinal, nonmetric MDS was carried out using Bray-Curtis distances by means of the majorization approach (Borg, Groenen, & Mair, 2012) using R package ‘smacof’ 1.9-6 (De Leeuw & Mair, 2011). MDS has found moderate success in several ecological province delimitation studies (Hale, 2010). However, it has been shown to put dramatic weight on species with high presence

counts. In this dataset, for example, there was a massive offset in two latitudinal bands due to relatively high counts of a handful of species (*e.g. Nassarius nigellus*) present almost exclusively in those two bands (supplementary figure 1). As a result, results based on this method are not presented here, and instead CCA and PCoA are used in conjunction with each other.

These ordination approaches are largely visual – breaks thought to be significant are determined subjectively. To complement them, faunal turnover was calculated using three similarity coefficients for each pair of latitudinal bands. Jaccard's coefficient has been used in several ecological studies looking at provinces and the effects of the environment on communities (Valentine, 1966; Itsukushima & Shimatani, 2015) and is a simple way of assessing the number of species that are common to two samples. It is shown to be downward biased when sampling is incomplete (Alroy, 2015). The Chao index (Chao et al., 2005) was developed as an improvement to Jaccard's that accounts for sample size differences in abundance data. The Forbes index (Forbes, 1907; Alroy, 2015) was shown to be more robust than the Jaccard index when sampling is uneven or incomplete. Pairs with a similarity of greater than 0.85 were considered to reflect very low turnover (almost homogenous) and could be considered part of the same biological province (Valentine, 1966), and regions of relatively high turnover could be considered transition regions or province boundaries (Briggs, 1995; CSIRO, 1996). Visual checks on correspondence analysis breaks were checked against faunal turnover patterns.

Hierarchical clustering was carried out using distances on each transform calculated using the Forbes index. Euclidean distances are not appropriate for ecological data (Legendre & Gallagher, 2001), Bray-Curtis distances will also be used for a comparison. Multiscale bootstrap resampling (10,000 bootstrap replications per distance matrix) was carried out on each cluster produced to calculate *p*-values for each node and identify strongly supported clusters using the R-package pvclust 2.0 (Shimodaira, 2004; Suzuki & Shimodaira, 2015).

Strongly supported clusters can indicate potential provinces that should be visually recognisable in an ordination plot. Endemism was calculated for each cluster identified in this way, and defined as the percentage of species present exclusively in that cluster (Engle & Summers, 2000; Hale, 2010). A standard cluster analysis was carried out using the `hclust` function in R.

Based on ranges calculated by using the ‘range-through’ method, the number of range endpoints was calculated for each latitudinal band. The locations of high endpoint clustering were checked against the sets of clusters suggested by each of the other methods, with range endpoints shown as indicators of provincial boundaries (Roy et al., 1994; MacPherson, 2003).

Abiotic variables

The minimum and maximum values over a yearly period for temperature and salinity were elucidated for each latitudinal band in order to identify upper and lower constraints on species distributions (Segura et al., 2016). Summer and winter means were also calculated for each band to account for seasonality (Condie & Dunn, 2006). Dissolved oxygen, nitrate and phosphate content were taken as a yearly average and a yearly maximum. Depth was not considered a factor in this analysis, due to all organisms being found within 0.01°E of the coastline, and so surface values (0-50m water depth) were used for each variable.

Each latitudinal band was assigned to a temperature category after Bousfield and Thomas (1975) using summer mean temperatures: <12°C, 12-15°C, 15-18°C and >18°C. Temperatures above 18°C were split by using the 75th percentile, 26°C, resulting in a total of five categories. Salinity across the whole extent was considered to be euhaline (30-49‰) in the Venice system (Anon, 1958). Thus, the median and 25th and 75th quartiles were used to assign the data into four salinity categories.

To investigate the effect of these variables on the ecology, a distance-based redundancy analysis (dbRDA) was carried out, using both Forbes dissimilarity

and Bray-Curtis distances. Redundancy analyses traditionally show the effects of “short gradients” (Van Den Wollenberg, 1977; Legendre & Gallagher, 2001), although this is not shown in more recent research (Legendre & Legendre, 2012; Oksanen et al., 2017), with dbRDA allowing for use of a multivariate distance measure and partitioning into a complex model (Legendre & Anderson, 1999; McArdle & Anderson, 2001). The analysis was carried out using temperature, salinity, dissolved oxygen, nitrate content and phosphate content in the water. These analyses show the interaction of each variable with the ordination results, and help identify which variables may be resulting in large biogeographical breaks.

The dbRDA was implemented using functions in the R-package ‘vegan’ 2.4-3 (Cajo, 1986; Oksanen et al., 2017).

Results

Latitudinal gradients

There are strong latitudinal gradients in Australian molluscan diversity in each treatment of the dataset (Figure 4), with a flat profile in tropical waters and a gentle decline towards the south resulting in a richness drop of > 60% across 33 degrees of latitude. Maximum diversity occurs around 23°S (the location of the Tropic of Capricorn).

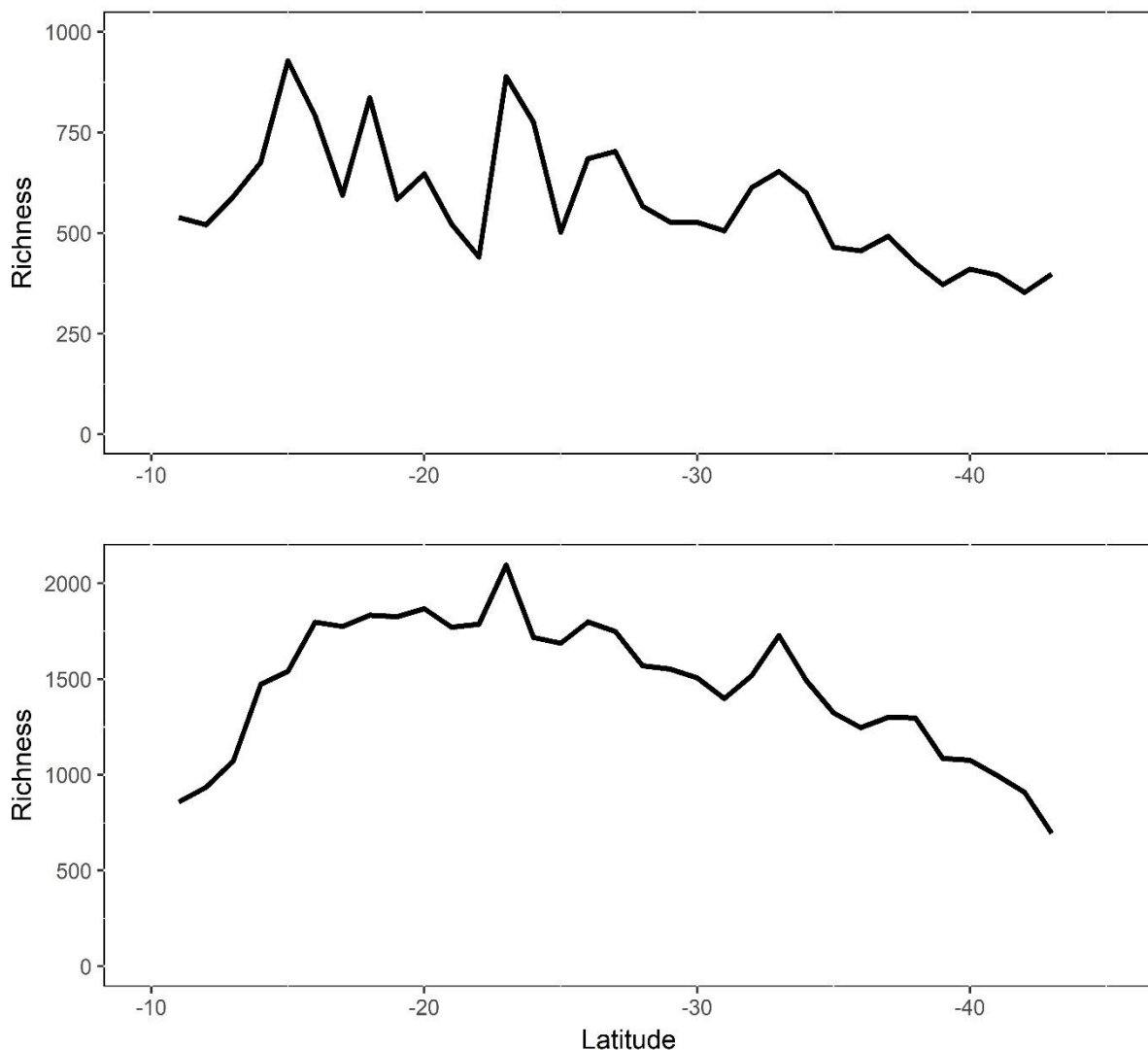


Figure 4: Latitudinal richness curves for marine bivalves and gastropods on the east coast of Australia. Patterns generated by two methods are shown: subsampling and extrapolating richness using the Chao2 estimator (A) and estimating species in a latitudinal band using range limits (B). Subsamples of 350 randomly pulled records were extrapolated. This process was repeated 1000 times for (A). Range limits were calculated using range extents from the dataset, based on museum collection records. An organism was included in a band if its range crossed that band for (B).

The ‘range through’ method indicates a much flatter northern gradient, with a drop-off towards the northern tip of Australia. As a result, the trend in richness is not as strong with respect to latitude (Spearman’s rank correlation ρ - 0.416, p -value 0.017) compared to the extrapolation method (Spearman’s rank correlation ρ -0.665, p -value <0.001 for Chao 2). All three extrapolation methods show a very similar result, with a noisier northern tail lacking the severe dropoff in the far north.

The extrapolated latitudinal gradient positively correlates with temperature and negatively correlates with dissolved oxygen content. However, salinity is not significantly correlated (Table 1). Median range size showed a moderate correlation with extrapolated richness but not with richness based on range sizes. Modal range size showed no correlation with richness.

Table 1: Spearman’s rank correlations between variables considered in this study and species richness in different latitudinal belts, calculated in two different ways.

Variable	Extrapolated (Chao1) Correlation coefficient	richness p -value	Species range richness	
			Correlation coefficient	p -value
Mean summer sea-surface temperature (°C)	0.645***	0.000074	0.431*	0.013
Yearly range in sea-surface temperature (°C)	0.351*	0.045	0.451**	0.0090
Dissolved oxygen content ($\mu\text{mol kg}^{-1}$)	-0.683***	0.000020	-0.517**	0.0024
Salinity (‰)	-0.203	0.25	0.177	0.32
Median range size (°)	0.491**	0.0037	0.284	0.11
Modal range size (°)	0.280	0.115	0.329	0.061

Bolded values are statistically significant. * $p < 0.05$, ** $p < 0.01$, *** $p < 0.001$

Biogeographic structure

Where there are multiple distance metrics used, only Forbes distances are shown for each analysis due to space restraints. The conclusions drawn from the data are identical for each transform, and the graphs and variances explained in each plot is incredibly similar. On request, the author can provide plots generated from Bray-Curtis or other distance metrics.

In the CCA and PCoA plots, distance between points is relative to dissimilarity in communities between the latitudinal bands: the null hypothesis is a random distribution. In the CCA plot, latitude exerts a very strong control, with an almost perfect ordering of points (Figure 5). In a ‘range-through’ data transformation (Figure 5b), the points do not separate into obvious clusters, with a continuous curve running through all points. Contrary to this, there is a strong clustering of points -38 through -43 in the ‘presence counts’ plot (Figure 5a), representing the whole of Tasmania and the Victoria coastline.

PCoA plots are very similar to the CCA plot with respect to the layout of points, with a significant break at 38°S for ‘presence counts’ (Figure 6a) and many small breaks for the ‘range-through’ analysis (Figure 6b), with a major break at 38 and 40°S. The variance explained by each axis was much higher for this analysis, with axis 1 explaining over 65%. Both CCA and PCoA retain the major break at 38°S for the ‘presence counts’ methodology when plotting axes two and three together.

Appendix Table 2 shows the variance explained by the major three axes produced by each ordination method and data treatment.

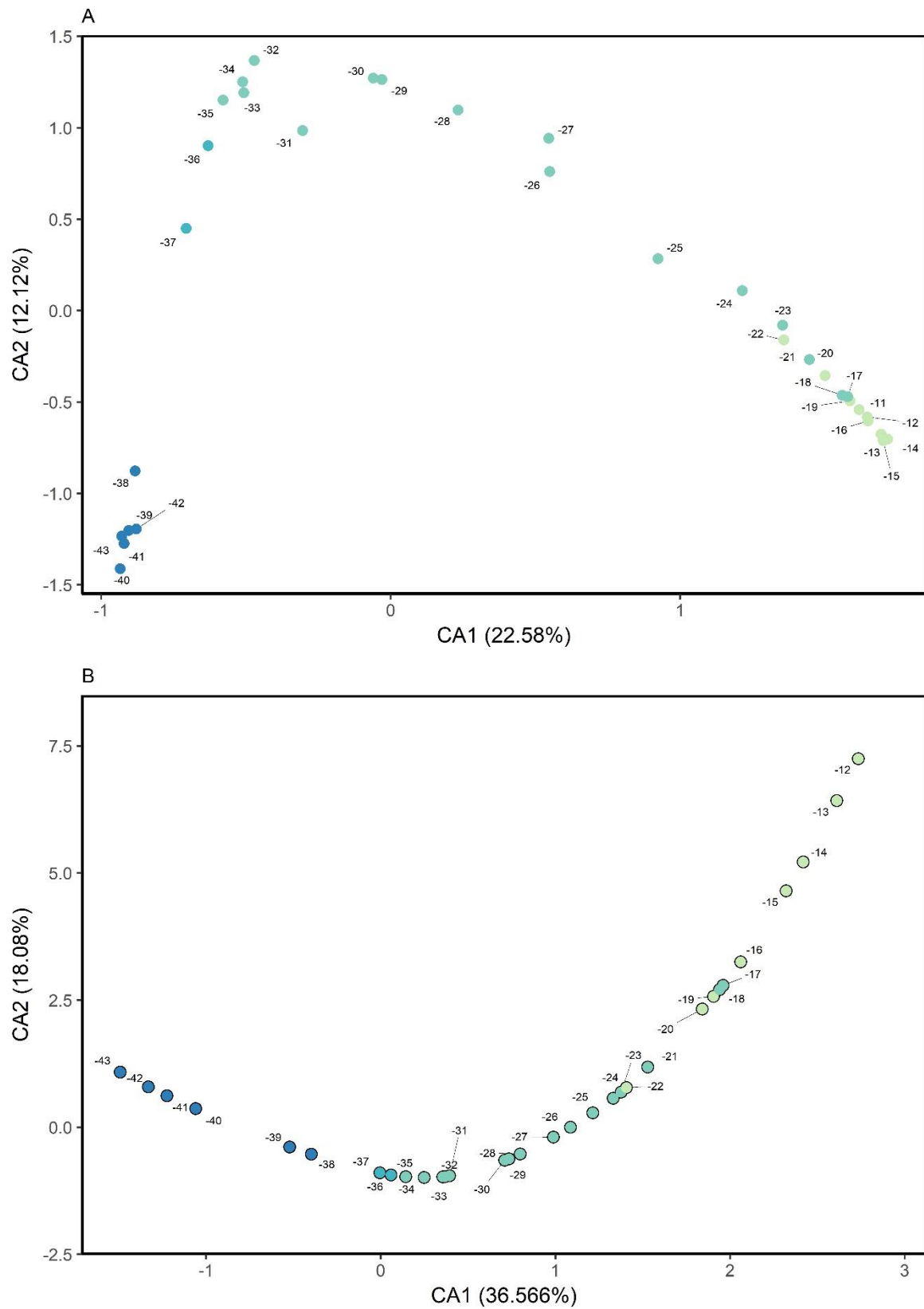


Figure 5: Results for a constrained correspondence analysis of occurrence data for east coast Australian molluscs. Two different transforms are shown, using presence counts (A) and estimating presences based on species ranges (B). Points are coloured according to their temperature category; 12-15°C (dark blue), 15-18°C (blue), 18-26°C (blue-green) and 26+°C (yellow-green). Temperature categories after Bousfield and Thomas (1975), with 75th quartile used as an upper divide. Labels on the points show the latitude of the midpoint of each band (e.g., -40 contains occurrences between -40.5 and -39.5).

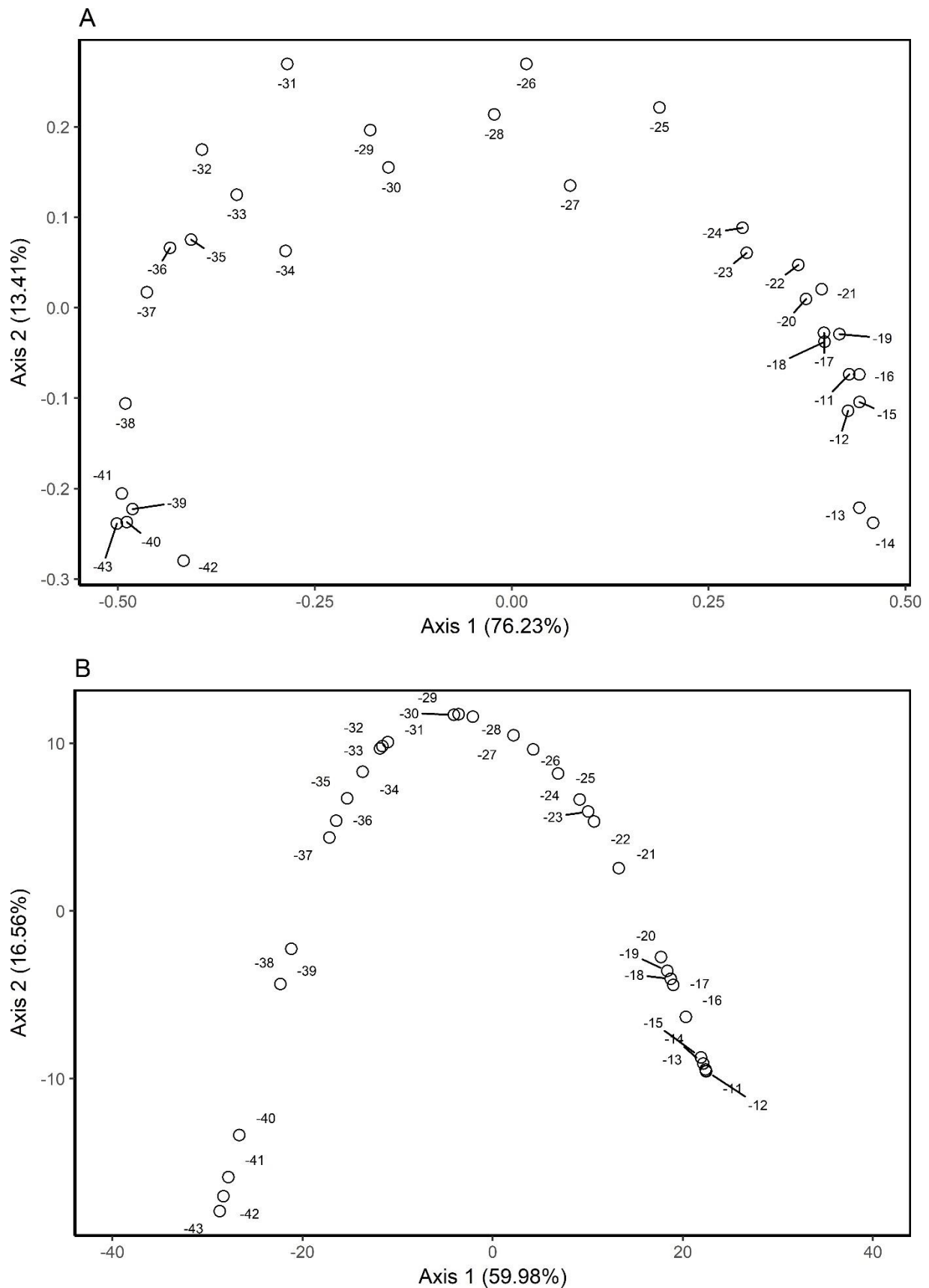


Figure 6: PCoA analyses of occurrence data for east coast Australian molluscs. (A) shows results based on the number of presences of each species, whereas (B) shows presence data extrapolated from known species ranges. Labels on the points show the latitudes of the midpoints of each band (e.g., -40 includes all occurrences between -40.5 and -39.5).

Faunal turnover was very stable across the whole extent, with no large variation in similarity between adjacent latitudinal bands (Figure 7). For Jaccard's coefficient only four pairs were above 85% similarity. Forbes and Chao indices show a high similarity south of -31° , but there are no distinct drops anywhere before and after this point, with turnover being very similar the whole way through. Variation in similarity between coefficients was very high, with peaks in similarity varying between measures significantly in places.

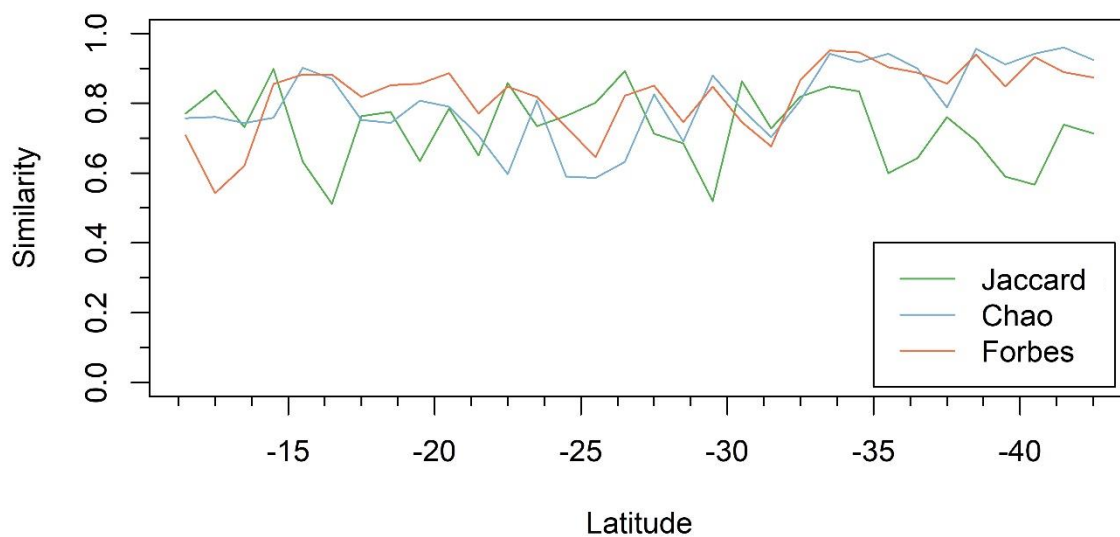


Figure 7: Similarity coefficients for adjacent latitudinal bands, utilising Jaccard coefficient, Chao index (Chao, Chazdon, Colwell, & Shen, 2005) and Forbes index (Alroy, 2015). Similarity is shown at the junctions between bands (e.g., the point at -20.5 represents the similarity between bands -21° and -22°).

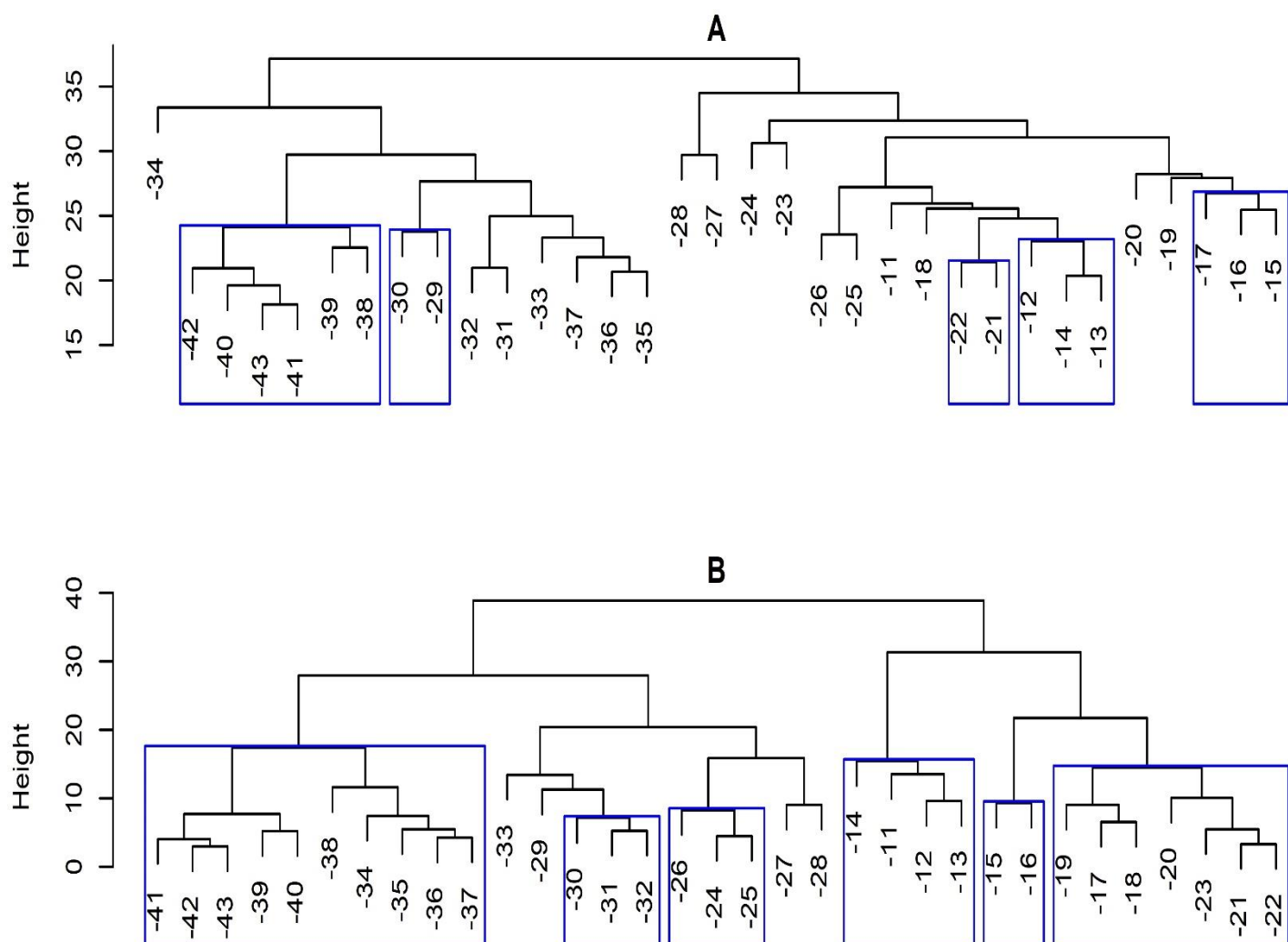


Figure 8: Results from hierarchal clustering of three data transformations; untransformed data using presence counts as a proxy for abundance (A) and a ‘range-through’ data matrix (B). Boxes denote clusters determined from multiscale bootstrap resampling (10000 replications) with a p-value of <0.01. Forbes dissimilarity used for each analysis.

‘Range-through’ analysis creates a multitude of smaller clusters ranging throughout the geographical extent, with only four bands falling outside any cluster (Figure 8b). Using the function `hclust` on Bray-Curtis distances shows a similar clustering of points between 38 and 43°S (Figure 9).

Range endpoints (Figure 10) suggest six boundaries. As no formal definition exists concerning what constitutes a boundary using this methodology, any band with an endpoint proportion of 0.2 or higher was used. No boundary was defined by an endpoint proportion of 0.4 or higher.

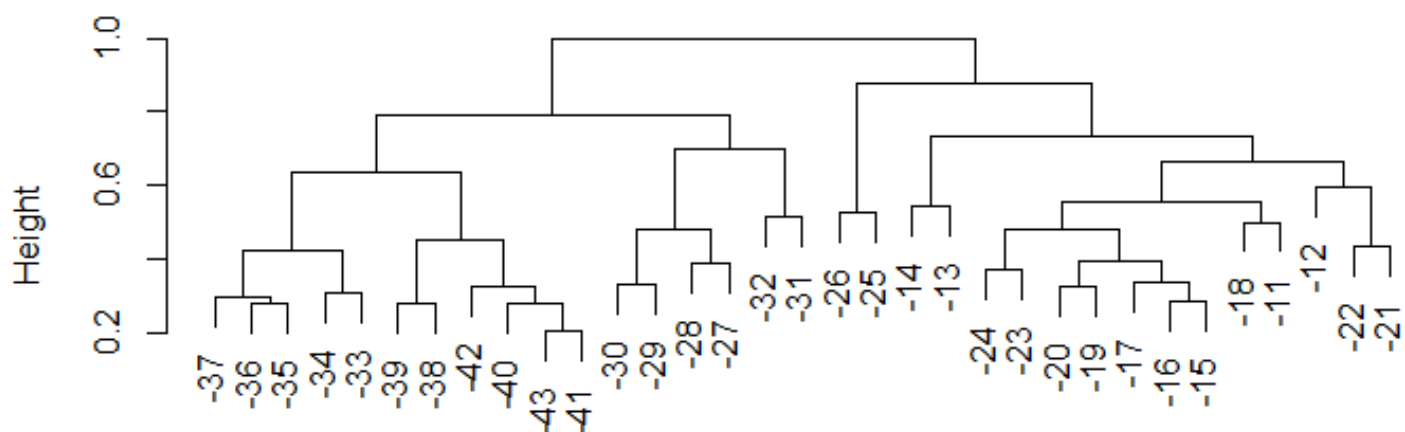


Figure 9: Results from hierarchal clustering of presence counts for molluscs on the east coast of Australia. Numbers of the tips denote the mid-latitude (degrees south) of the band (-38, for example, would contain all points between 37.5 and 38.5°S). Bray-Curtis distances used.

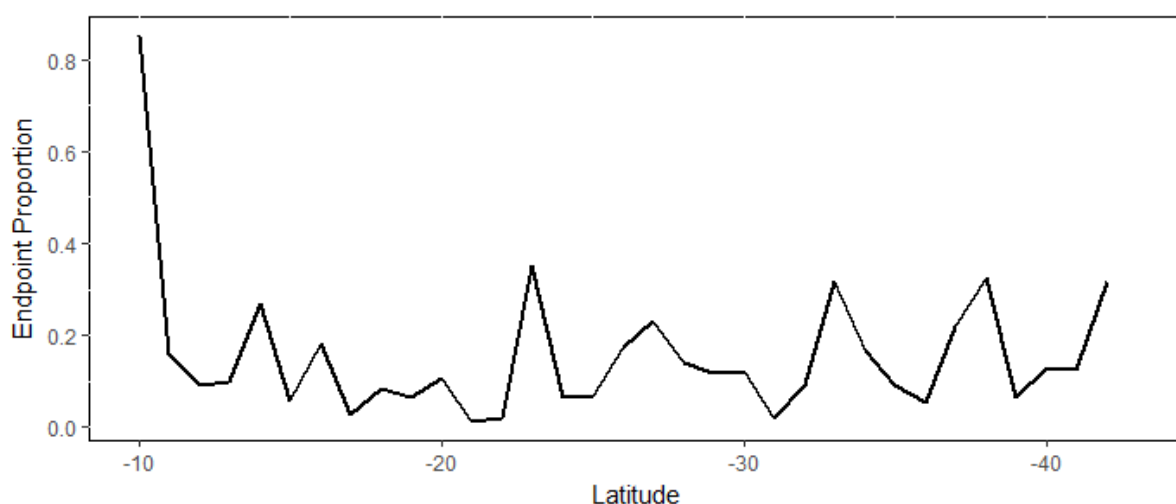


Figure 10: Latitudinal distribution of range endpoints of bivalve and gastropod species on the east coast of Australia, binned by degree of latitude. Points represent the proportions of species that have range ends in respective bins, meaning the proportion of all species that have ranges falling in particular bins.

Cluster endemism is summarised in Table 2. Southern clusters had high endemism, with >20% endemic species in both treatments of the data, whereas northern clusters included <5%. Cluster size across latitudes was significantly correlated with endemism measures for northern clusters based both treatments

and with all provinces included (Spearman's rank correlation $\rho = 0.728$, p -value = 0.01).

Table 2: Endemism, the number of species found in an area and nowhere else along the coastline, for each cluster, produced by multiscale bootstrap resampling. Only clusters with four or more consecutive latitudinal bands are shown.

Methodology	Cluster Extent	Endemism
Presence counts	38-43°S	20.8%
Range-through	34-43°S	21.0%
	17-23°S	3.15%
	11-14°S	2.84%

Relation of biogeographic structure to abiotic variables

Temperature (Figure 11a) had a strong relationship with CCA axis 1 for both data treatments, suggesting that axis 1 is a latitudinal or temperature control. However, there were no cases in which breaks between temperature categories corresponded with breaks suggested by the presence-count analysis (Figure 4a) or the ‘range-through’ analysis (Figure 4b). Cluster boundaries found by multiscale bootstrap resampling also did not map with changes in temperature category.

Dissolved oxygen (Figure 11b) tracked the inverse of temperature, with a steady increase south of 23°S totalling 50%. North of this, there is also an increase moving southwards, excepting for a slight decrease from 210 to 190 $\mu\text{mol kg}^{-1}$ between 20°S and 23°S. Maximum dissolved oxygen content occurs at 44°S and minimum dissolved oxygen content occurs at 16°S.

Salinity was relatively stable across the entire geographic range, with the entire extent falling into one salinity category - “euhaline” (Anon, 1958). However, fluctuations do occur (Figure 11c). In southern latitudes, salinity varies by less than 0.5‰. However, between 28°S and 29°S there is a large drop of approximately 3‰ before a steady increase of 5‰, with the minimum salinity

occurring at 27°S and maximum salinity occurring at 32.5°S. Immediately afterwards, the minimum salinity is a 0.5° thick band at the same level before the drop. After maximum salinity, there is a steady decrease of 2‰ towards the southern extent. Salinity is steady at areas of province boundaries identified.

Nitrate content is almost uniformly low in a yearly average, with a large peak towards very southern latitudes (Figure 12a). Maximum nitrate content is at the southern extent of the dataset, nitrate content north of 40°S is uniformly low. Phosphate content is uniformly very low across the whole dataset (Figure 12b), with a maximum at the southern extent of the dataset. Smaller phosphate peaks are present at the north end and in the mid-latitude bands.

For dbRDA, the projection of a point onto a response variable approximates the points value for that variable. Lines for response variables are shown here in the positive direction only. Nitrate (NTR) content shows a strong effect on very southerly latitudes (figure 13), as expected from the large peaks in southern latitudes. The tight clustering of southern latitudes seen in CCA and PCoA aren't as pronounced in this analysis, with a much looser structure and an apparent break between Victoria and Tasmania present as well as a break at 37/38°S. Dissolved oxygen content (DOX) and temperature (TC) correlated strongly with axis 1 in a single direction, roughly equivalent to latitude. There is no significant change along any explanatory variable for break points in the dataset. ANOVA permutation test results show both temperature and nitrate content as significant (p -value<0.01 for both variables) drivers in the ecological structure, with other variables not significant. Salinity is not shown in the results here due to the incredibly small effect it has on the structure. When included, it has no significant effect on community structure across the dataset (p -value=0.125).

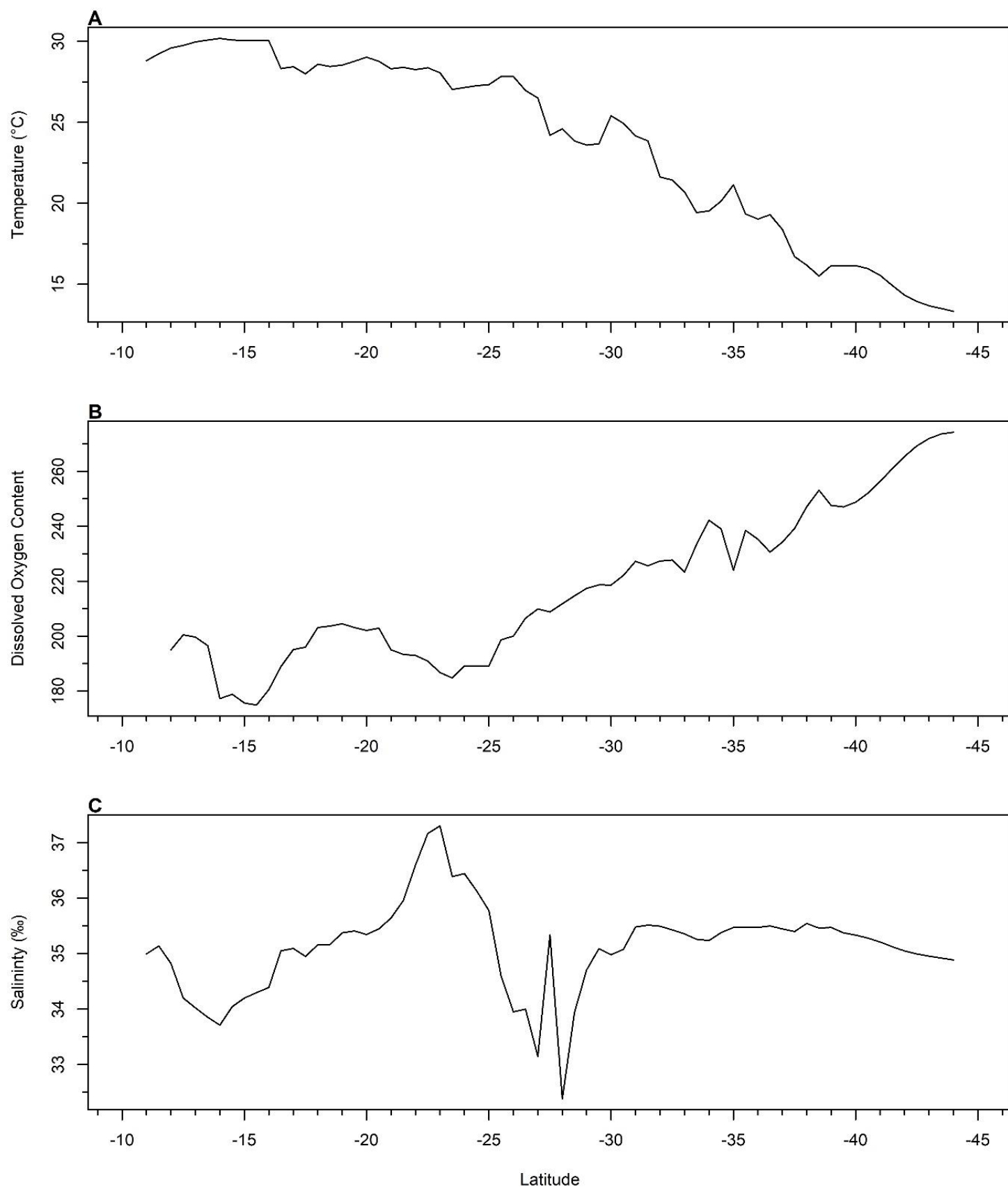


Figure 11: Abiotic variables by latitudinal belt for the east coast of Australia: temperature (A), dissolved oxygen content (B – expressed as $\mu\text{mol kg}^{-1}$) and salinity (C). Values are taken every 0.5° of latitude as summer means for the period 2007-2012.

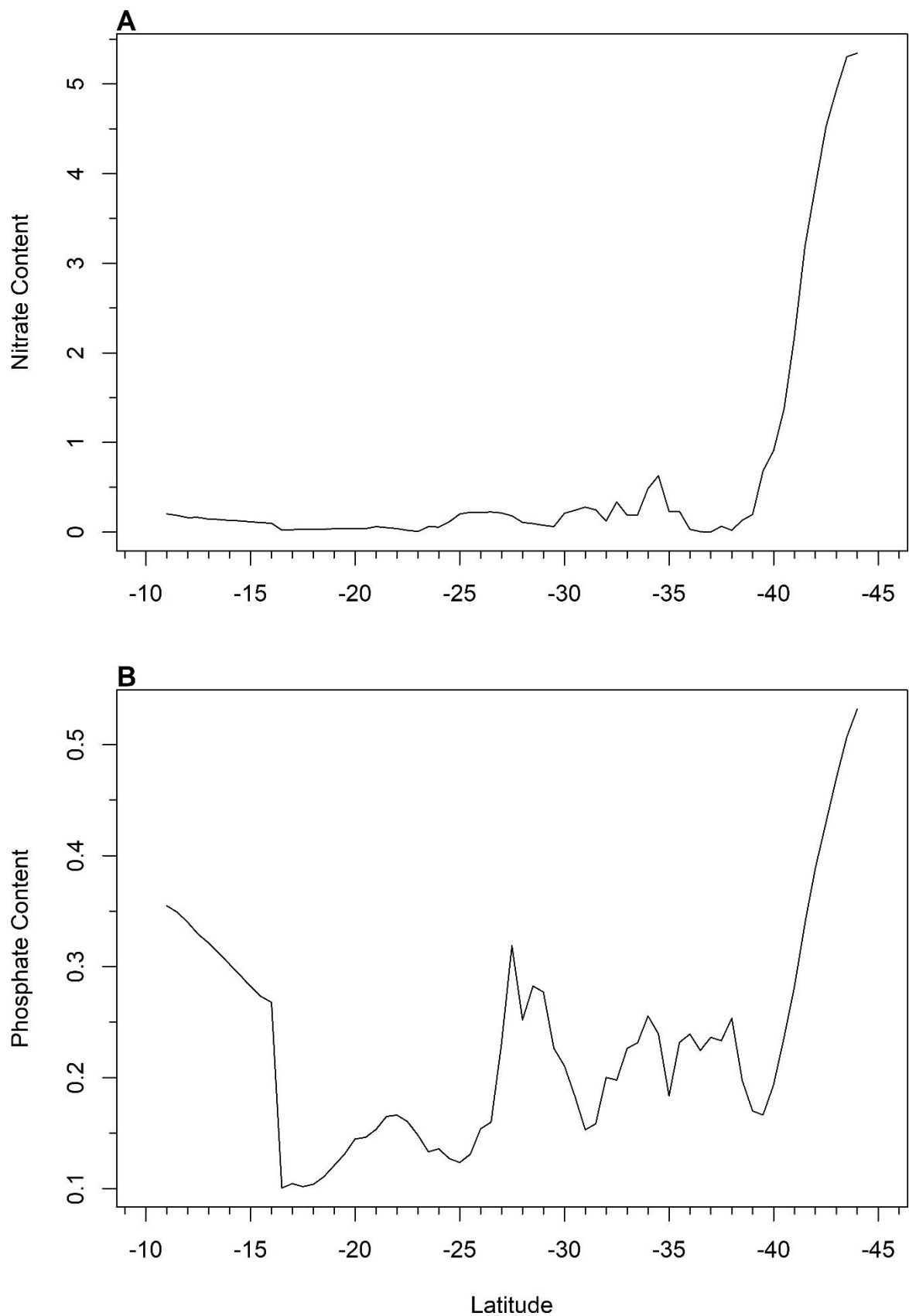


Figure 12: Abiotic variables expressing nutrient content on the east coast of Australia; nitrate (A) and phosphate (B). Values are expressed as $\mu\text{mol/kg}$, and taken every 0.5° of latitude. Values are summer means for the years 2007-2012.

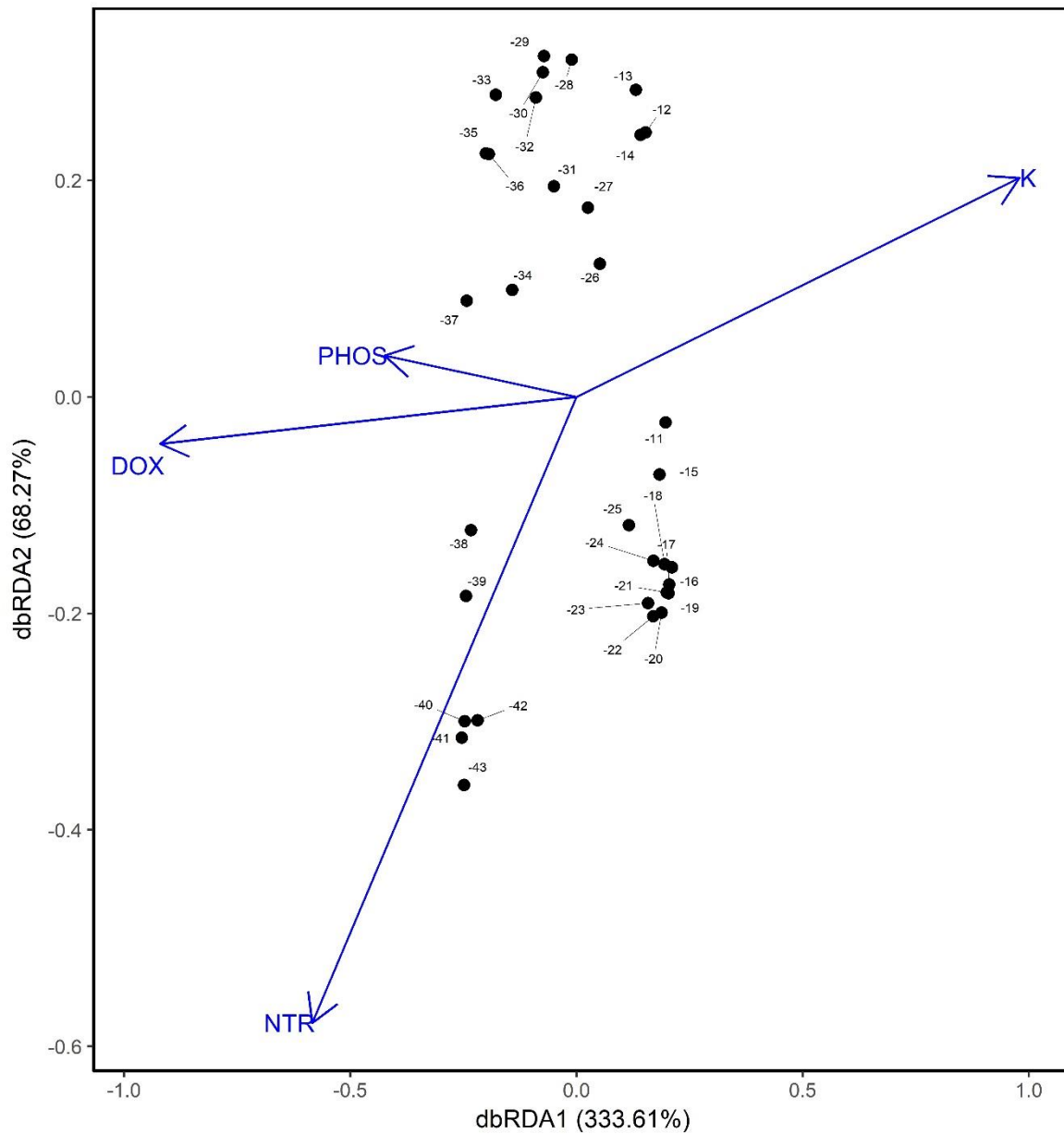


Figure 13: Results for a distance based redundancy analysis on presence data for molluscs on the east coast of Australia. Points mark the location of latitudinal bands, with the annotation representing the mid-point of that band (*e.g.* -38 would represent the band 37.5-38.5°S). Arrows represent the effects of abiotic variables: DOX - dissolved oxygen content, K - temperature, PHOS – phosphate content, NTR – nitrate content. The dbRDA was implemented as in McArdle and Anderson (2001) using the formula; $(species\ presence\ table) \sim TC + DOX + NTR + PHOS$.

Discussion

Interpretation of results

Molluscs along the east coast of Australia exhibit a strong latitudinal gradient, and this gradient is strongly correlated with temperature and dissolved oxygen content. With both yearly range in temperature and mean temperature being proxies for energy availability in the system, these results agree with many previous studies linking this key factor to the nature of the latitudinal gradient (Roy et al., 1998). Dissolved oxygen content is the inverse of temperature in this study and therefore shows the opposite pattern. It should be noted that the data here are autocorrelated and therefore these relationships do not offer an explanation for the latitudinal trend. Rather, they show a relationship of interest to the link between biodiversity and ecological structure for the east coast of Australia.

The latitudinal diversity curve, however, did not display any stepwise patterns that would be indicative of provincial boundaries, regardless of the data treatment. Roy et al. (1994) and others (Roy et al., 1998; Roy et al., 2000; Mittelbach et al., 2007) did present stepwise curves for other regions. Instead, the Australian gradient is much smoother. The Jaccard index indicates that there are no clear province boundaries along any part of the coastline, although loose southern clusters are suggested by the Forbes index and Chao index. The Jaccard coefficient, which forms the basis for the provinces in Valentine (1966), identified only four latitudinal pairs with $> 85\%$ similarity, and assigns no distinct provinces if his method is emulated. The other two similarity coefficients have not been used for defining province boundaries. However, they depict a southern cluster (south of 31°S) in which pairs have $> 85\%$ similarity.

These methodologies are likely to be old-fashioned and inaccurate compared to modern ordination methods. For example, the southern cluster's boundaries do not correspond with the distinct breaks produced by other methods.

CCA analysis presents a visual break between 37/38°S, which corresponds to a Victoria/Tasmania province, and no other visible break in the sequence (Figure 4a). PCoA confirms these results: the same break is present when comparing axes 1 to 2 and axes 2 to 3. In a cluster analysis this band is again supported, with bootstrap resampling demonstrating a very significant cluster encompassing the same geographical area.

The endemism analysis indicates that a southern cluster is highly probable, with candidate province boundaries at both 34 and 38°S having endemism rates of > 20%.

No relationship between the majority of abiotic variables tested and any provincial boundary was found in any analysis. This is unusual because temperature is often cited as a key driver in the formation of provincial boundaries (Valentine, 1966) due to its strong control on survival during juvenile stages (Verween, Vincx, & Degraer, 2007). Low salinity has been related to reproductive success in gastropods (Segura et al., 2016) and to the absorption of amino acids from seawater (Deaton, Derby, Subhedar, & Greenberg, 1989). However, molluscs generally have salinity tolerance ranges that fall within the whole salinity range of the current data (Nell & Gibbs, 1986; Verween et al., 2007).

The only variable that seemed to relate to provincial structure was nitrate content – this variable is incredibly high in southern waters due to the influx of nutrient rich waters from the Antarctic (Cresswell, 2000; Ridgway, 2007; Harvey et al., 2015). When included in the dbRDA, a province boundary was still present at 37/38°S.

It can be inferred, therefore, that the Victoria/Tasmania cluster relates to the structure of the East Australian Current, with reported turbulence greatest at the south-eastern corner of Australia (Huyer *et al.*, 1988) and the current being directly related to nutrient influx in the system (Cresswell, 2000). No fine-scale

current data were tested in this study, and they should be incorporated before any connection is argued for.

Data treatments were shown to have a significant effect on hypothesised province structure, as expected. However, the biggest difference was between using the existing dataset and extrapolating presences based on species ranges ('range-through'). Species ranges have been used to define LBGs and range endpoints in a multitude of related studies (Roy et al., 1994, 1996; Roy et al., 1998; MacPherson, 2003; Valdovinos et al., 2003). However, this methodology appears to oversimplify the system by failing to account for small-scale changes in species distributions and interactions between sets of species. This behaviour is clearly seen here, with 'range-through' analyses giving a much higher weighting to latitude alone as a provincial driver – perfectly ordering points in all ordination methods to match a latitudinal trend (Figure 4b, 6b, 9c). They also elucidate a much higher number of provinces than any other methodology in a cluster analysis (Figure 7c), with all clusters all exactly matching with peaks in endpoint proportions.

This is a circular argument – using published ranges to extrapolate species presences will naturally create clusters around endpoints that may not have any biological meaning. The absence and relative presences of organisms within their ranges are much more telling than the ranges themselves, and so provinces elucidated using some form of abundance data are likely to give more accurate results with respect to geographic similarity. For Australia, range endpoint clustering is a poor methodology for defining boundaries, as suggested by the fact that no range endpoint cluster appears in any ordination as a biogeographical break.

Consequences for Australian biogeography

These results demonstrate that there is a provincial divide present in the south east of Australia, with a province containing Tasmania and Victoria. This location is similar to traditional splits in this region (Bennett & Pope, 1952, 1960), but, unlike more recent analyses (e.g. CSIRO, 1996), does not place Tasmania as a separate region. However, the lack of any clear divide along most of the eastern Australian coastline goes against every published bioregionalization analysis over the past 60 years, with even the simplest binary divides (e.g. Wilson & Gillett, 1974) splitting this coastline. This single provincial divide is supported by every analysis presented here.

The provincial scale shown here is of considerable size when compared to provinces determined both in Australia and worldwide. Coastal USA, for example, has only one province of the size shown on the mainland here (Roy et al., 1994), with many smaller provinces instead being present (Valentine, 1966). Even when range through analyses are taken into account, in a similar way to the methodology employed by Roy et al. (1994), the provinces elucidated are far smaller in scale.

The differences between provinces found in the present study and those found in previous works are likely due to disparity between each study. Taxonomic group never remained constant, and full coastal collections were not available at any time, instead relying on previous published works (Bennett & Pope, 1952, 1960) or on distribution data (Wilson & Gillett, 1974; CSIRO, 1996). The biogeographic split shown here has been supported in the past (Bennett & Pope, 1960; CSIRO, 1996), and the division is supported in global ocean province maps (Spalding et al., 2007; Oliver & Irwin, 2008).

There are, however, no clear relations between the boundary and any abiotic variable, including temperature, salinity and dissolved oxygen content. Instead, the boundary appears to correlate with recorded extents of the East Australian Current (Huyer, Smith, Stabeno, Church, & White, 1988) and with the

coastline's physical extent. The stretch of the coastline at the boundary is far greater than that of any other latitudinal band due to Victoria having an east-west coastline instead of a north-south coastline. In addition, as an island, Tasmania has two north-south orientated coastlines influenced by different oceanic processes.

Tasmania as a system

All large-scale studies have shown no break between the eastern and western halves of the island (Bennett & Pope, 1960; CSIRO, 1996) and instead show potential differences between the water masses to the north and south (CSIRO, 1996). It is becoming more and more recognised that there may indeed be differences between all aspects of the island in terms of ecology. Recent phylogeographical studies show at least four distinct ecological zones on the island (York, Blacket, & Appleton, 2008; Teske et al., 2017), with the largest changes between east and west, and within the Bass Strait (Teske et al., 2017).

The most likely explanation for these changes are the differing current regimes in eastern and western Tasmania. On the east coast, the East Australian Current brings warm, nutrient poor water from the north down to Tasmania (Ridgway, 2007). The west coast is affected by the Zeehan Current from the west, which during winter wraps around and penetrates towards the north against the East Australian Current (Cresswell, 2000; Ridgway, 2007), during summer the East Australian Current reaches the southern tip of Tasmania. These current systems provide vastly different water types and nutrient levels to both halves of the island, with strong seasonal variation (Harris, Nilsson, Clementson, & Thomas, 1987). It is very likely this results in very different ecological structure and in differing relationships with the Victoria coastline.

These relationships may also have changed temporally. The intensity of the East Australian Current, and its seasonal variation, has been noted to have increased significantly with ocean warming and acidification (Johnson et al.,

2011; Rodolfo-Metalpa et al., 2011). These changes may also be affecting the ecological make up of eastern Tasmania more so than western Tasmania, with molluscan communities being affected by not only the changing currents but also directly by the water changes (Nell & Gibbs, 1986; Rodolfo-Metalpa et al., 2011).

In preliminary analyses done by the author, western Tasmania shows a much higher similarity with Victoria than eastern Tasmania when divided simply into east and west coasts. A full analysis of this type is beyond the scope of this study - there is a distinct lack of data for the western coastline in this dataset. Full ecological sampling would be needed to elucidate any real biological differences between these areas. Of particular interest would be small scale changes in southern Tasmania ecosystems due to changes in nutrient availability on a yearly basis.

Conclusions

Based on the aims proposed earlier, the following was achieved during this project:

1. To construct a database of molluscan occurrences and ranges for the east coast of Australia, and use it to investigate the key drivers of coastal molluscan latitudinal diversity in terms of abiotic variables.

The final dataset numbered 168,601 occurrences for the entire stretch of eastern Australia, containing 3796 species and 260 families. Molluscs along eastern Australia exhibit a strong latitudinal diversity gradient that is strikingly different in structure from that of both North American coastlines, with a much smoother profile and no stepwise pattern matching with existing biological provinces. Species range sizes show no correlation with this gradient, but temperature (a proxy for energy availability) and dissolved oxygen content were strongly correlated with it. No other abiotic variable was found to be correlated with latitudinal richness.

2. To designate molluscan biogeographical provinces for the east coast of Australia by applying several statistical methods that have been used in the literature to assign provincial boundaries.

Assigning provinces to Australia is extremely difficult, as each method depicted slightly different biogeographical structure. There was strong support for a across a few methods for a province including southern Victoria and Tasmania. However, the rest of mainland Australia remained difficult to split into distinct regions. Faunal turnover was relatively stable and high across the whole area, suggesting that distinct province boundaries may not exist for the mainland.

The ‘range-through’ method proved ineffective when looking for province boundaries, and no relation could be found between the suggested provinces and the nature of the latitudinal gradient. Modern ecological sampling would provide much better resolution when considering ecological change across the area.

3. To investigate what abiotic variables that define the start and end of each province, and to see if they are instead purely controlled by geographic features.

All suggested province boundaries were unrelated with previously defined temperature boundaries, salinity or dissolved oxygen content. Large-scale studies looking purely at these variables to define province boundaries may be inaccurate for this region. When nutrient content was included in the analysis, a strong southern effect was shown for nitrate content in the water. It is likely, therefore, that the boundary is nutrient driven, with conflicting current regimes separating nutrient-poor northern waters from nutrient-rich southern waters.

Recommendations for further study

The main outcome in this study was the lack of a consistent dataset for the full extent of the coastline in Australia, especially in terms of temporal coverage and availability of bona fide ecological abundance data. A future study would be

primarily based on ecological sampling, ideally covering all the coastline and collected within a season. A comparison between this collection and robust data collections from 1960-1980 in the existing dataset could then be used to investigate how province boundaries have changed over time, as well as more accurately define them.

In addition, higher resolution sampling for eastern and western coastlines of Tasmania is needed to fully differentiate these coastlines, if indeed there is a major ecological difference, and settle debate over the uniformity of Tasmanian molluscan ecology. For the next stage of this study, the data should be split into degree squares, and not just by latitude, for the southern regions.

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Appendices

Table 1: List of contributing institutions, with numbers of records contributed. Data were taken from OZCAM and filtered (see text); only the final dataset is shown here.

Institution	Records Contributed
Australian Museum	97826
Museum and Art Gallery of the Northern Territory	628
National Museum of Victoria	24343
Queensland Museum	602
Queen Victoria Museum and Art Gallery	1482
Tasmanian Museum and Art Gallery	19127
Western Australian Museum	73

Table 2: Variance explained (%) by major axes in the Constrained Correspondence Analysis (CCA) and Principal Coordinates Analysis (PCoA) for each data treatment.

Transform	Variance explained by axis		
	1	2	3
Presence counts (CCA)	22.5	12.1	6.69
Range-through (CCA)	36.6	18.1	9.95
Presence counts (PCoA)	76.2	13.4	6.52
Range-through (PCoA)	60.0	16.6	7.98

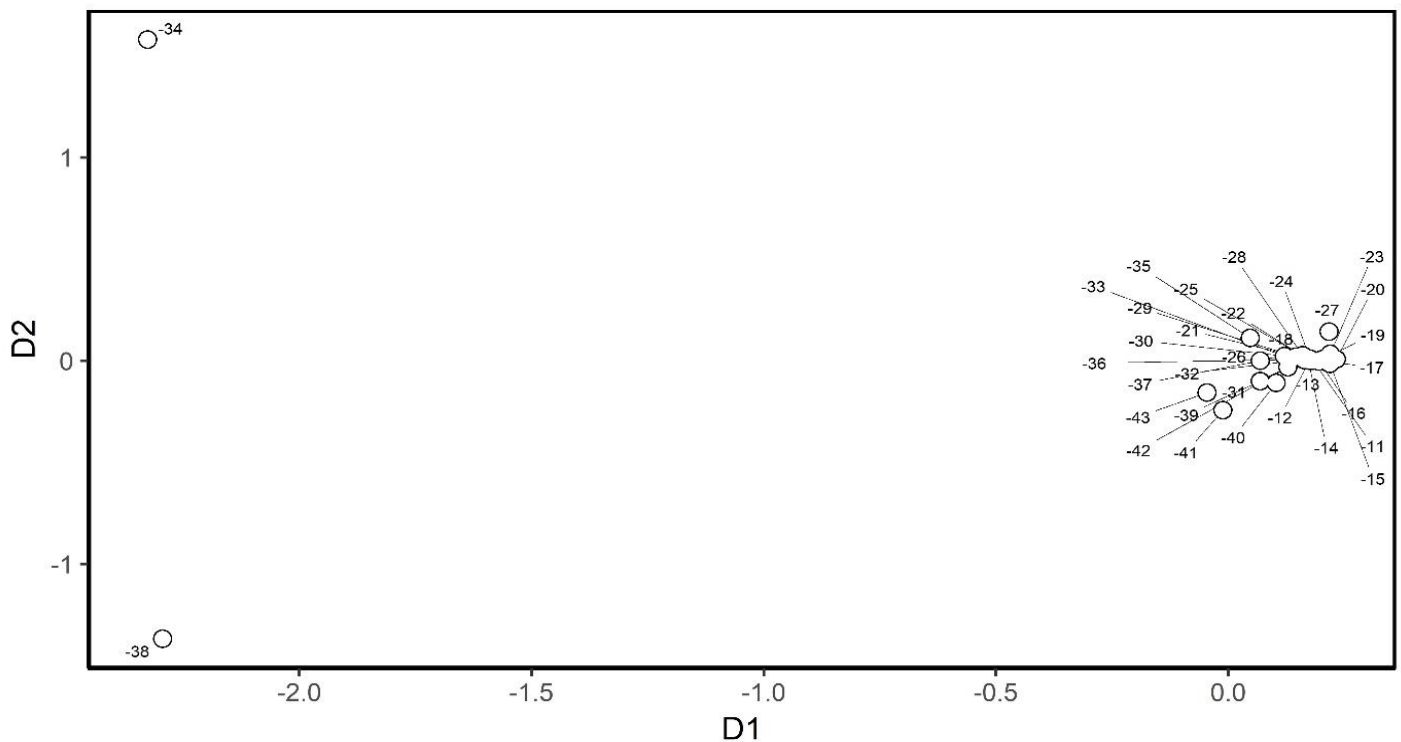


Figure 1: Results for ordinal, nonmetric multidimensional scaling of occurrence data for east coast Australian molluscs using two dimensions. The results here are based on a matrix of presence counts, with each point shown as the midpoint of each latitudinal band (-39 would be all points between -38.5 and -39.5, for example). -34 and -38 have very high counts of a small number of species, relative to the whole dataset, and thus sit far out from the main cluster of points.

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